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Estimation of caffeine intake from analysis of caffeine metabolites in wastewater

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HIGHLIGHTS
• Analysis of caffeine and metabolites in wastewater in ten European cities
• Comparison of metabolic profiles in wastewater and in human urine
• Selection of a suitable biomarker for assessing population level caffeine consumption
• Use of 1,7-dimethyluric acid for caffeine consumption back-calculation
• Comparison of caffeine intake from wastewater vs coffee trade by country per capita

ABSTRACT
Caffeine metabolites in wastewater were investigated as potential biomarkers for assessing caffeine intake in a population. The main human urinary metabolites of caffeine were measured in the urban wastewater of ten European cities and the metabolic profiles in wastewater were compared with the human urinary excretion profile. A good match was found for 1,7-dimethyluric acid, an exclusive caffeine metabolite, suggesting that might be
a suitable biomarker in wastewater for assessing population-level caffeine consumption. A correction factor was
developed considering the percentage of excretion of this metabolite in humans, according to published pharma-
cokineti c studies. Daily caffeine intake estimated from wastewater analysis was compared with the average daily
intake calculated from the average amount of coffee consumed by country per capita. Good agreement was found in
some cities but further information is needed to standardize this approach. Wastewater analysis proved useful to
providing additional local information on caffeine use.

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1. Introduction

History suggests that caffeine has been used, in one form or another, since ancient times. In 2737 BCE a Chinese Emperor used the leaves from a nearby bush to prepare a tea (Arab and Blumberg, 2008; Heckman et al., 2010). An old legend dates the use of coffee to the 9th century in the southern tip of the Arabian Peninsula when a shepherd noted euphoria and stimulating effects on his goats caused by eating wild coffee berries. He then decided to try them himself. Coffee later crossed to Africa and in the 1600s reached Europe becoming, over the centuries, the most commonly consumed beverage worldwide after water (Butt and Tauseef, 2011).

Caffeine is a naturally occurring alkaloid found in beans, leaves and fruits of ~60 plant species. The world’s main sources are coffee beans (Coffea arabica and Coffea robusta) and tea leaves (Camellia sinensis). It is also naturally found in kola nuts (Cola acuminata), cacao beans (Theobroma cacao), yerba mate (Flex paraguariensis) and guarana berries (Paullinia cupana). Most caffeine is consumed with beverages such as coffee, tea and soft drinks (including “energy drinks”), while products containing cocoa or chocolate, and medications such as some analgesic formulations and dietary supplements contribute small amounts to the diet (Heckman et al., 2010). Total daily intakes vary throughout the world although coffee usually contributes significantly more than other drinks to overall caffeine consumption (coffee 71%, soft drinks 16% and tea 12%), particularly among adults (Heckman et al., 2010; Mitchell et al., 2014). Carbonated soft drinks are the main source of caffeine for children (Mitchell et al., 2014).

Chocolate contains on average around 1.3% of theobromine, 0.75% of caffeine and theophylline in small amounts; cola nut between 2 and 3.5% of caffeine, theobromine (between 1 and 3.5%) and small amounts of theophylline, and tea leaves around 3% of caffeine (theophylline and theobromine in small amounts). This results in around 40–80 mg of caffeine per cup of tea (150 mL) while caffeine content in cocoa commercial products ranges from 2 to 7 mg (Barone and Roberts, 1996) and 5–20 mg/100 g in chocolate candy products. In soft drinks, variable levels of caffeine have been reported depending on the brand but the typical content is around 40 mg/360 ml (Chou and Bell, 2007). All these products contain relatively little caffeine compared to the average content of a coffee cup (60–150 mg/150 mL).

Caffeine is extensively metabolized by the human liver to form three major metabolites by demethylation: 3,7-dimethylxanthine (known as theobromine), 1,7-dimethylxanthine (paraxanthine) and 1,3-dimethylxanthine (theophylline). These are then broken down further in the liver by additional demethylation and oxidation and are excreted mostly in the urine (Heckman et al., 2010).

While there is no specific recommendation for human caffeine intake, it is considered that average consumption of approximately 300 mg/day is not associated with adverse health effects (Fitt et al., 2013; Higdon and Frei, 2006). However, data about caffeine intake in the population are scarce. Caffeine consumption is usually assessed by dietary surveys, but getting accurate information in this way presents many limitations. For instance, subjects may under-report their caffeine intake when food diaries are completed or information is missing about the strength, brand or amount of caffeine product they have consumed, which may greatly affect the intake. Another limitation is that in caffeine dietary surveys the subjects are usually asked about the consumption of certain beverages (mainly coffee and tea) but other products containing caffeine are not considered; for example, analgesics can contain as much as 200 mg caffeine per tablet (Derbyshire and Abdul, 2008). Another limitation for estimating the total caffeine intake is that the caffeine content of various drinks, food and dietary supplements is only known in some countries such as the USA (Fitt et al., 2013).

A complementary method would be to estimate consumption in the general population using the levels of caffeine and its metabolites measured in urban wastewater as biomarkers of intake. This approach, called wastewater-based epidemiology (WBE), has been mainly applied in the last decade for estimating illicit drug consumption (Baker et al., 2014; Ort et al., 2014; Thomas et al., 2012; Zuccato et al., 2008) and more recently has also been proposed for the quantitative measurement of lifestyle habits such as tobacco and alcohol use, exposure to environmental and food contaminants or factors related to health and illness in a community (Lopes et al., 2014; Reid et al., 2011; Rodriguez-Alvarez et al., 2015; Roussis et al., 2017; Thomas and Reid, 2011; Yang et al., 2015). The main advantage of WBE is that it provides objective, up-to-date information about the use of these substances in a population and can therefore complement current epidemiological methods.

In this study, the presence of caffeine and some selected metabolites was assessed in untreated wastewater in ten European cities. Levels in wastewater were compared with those measured in urine and with the human excretion profiles of caffeine reported in the literature in order to correlate the results from the different sources. 1,7-dimethyluric acid, an exclusive caffeine metabolite, was selected for estimating collective caffeine consumption. The reliability of this compound for caffeine back-calculation was evaluated by comparing the amounts measured by wastewater analysis with the average amount of coffee consumed in each country per capita.

2. Materials and methods

2.1. Chemicals and reagents

Caffeine (1,3,7-trimethylxanthine), paraxanthine and 1-methylxanthine were purchased from Sigma Aldrich (St. Louis, MO, USA); 1-methyluric acid, 1,7-dimethyluric acid 7-methylxanthine were purchased from Santa Cruz Biotechnology, Inc. (Santa Cruz, California, USA). Standard solutions at 1 mg/mL were prepared in methanol, except for 1-methylxanthine, 7-methylxanthine, paraxanthine and 1,7-dimethyluric acid which were prepared in methanol-water (50/50) at pH 8.5–10 (adjusted with 25% ammonia to enhance solubility). A mix of all compounds at 10 ng/μL was prepared in methanol and then diluted to 1.0, 0.1 and 0.01 ng/μL. Isotopically labeled compounds were caffeine-13C3 purchased from Sigma Aldrich and 1,7-dimethyluric acid d3 from Santa Cruz Biotechnology. Labeled internal solutions were prepared separately. Internal standard mixtures with 1 ng/μL of caffeine-13C3 and 10 ng/μL of 1,7-dimethyluric acid d3 were used as surrogates.

All solvents were of reagent grade or higher. Methanol for pesticide analysis and ammonium acetate were from Carlo Erba Reagents (Italy).
Ammonium hydroxide solution (25%) was acquired from Fluka (Buchs, Switzerland), LC-MS grade acetonitrile and hydrochloric acid (HCl 37%) were supplied by Riedel de Haen (Seelze, Germany). Water was purified using Milli-RO Plus 90 apparatus (Millipore, Molsheim, France). Solid-phase cartridges (3 mL Oasis HLB, 60 mg) and HPLC XTerra C18 column (3.5 μm, 1 mm × 100 mm) were obtained from Waters Corp., Milford, MA, USA.

2.2. Wastewater samples

24-Hour composite influent wastewater samples were collected from ten wastewater treatment plants (WWTP) in different European cities: Bristol (UK), Brussels (Belgium), Castellón (Spain), Copenhagen (Denmark), Lugano (Switzerland), Milan (Italy), Oslo (Norway), Porto (Portugal), Utrecht (Netherlands) and Zurich (Switzerland) (Table S2 in Supplemental Information-SI). Samples were collected daily for seven consecutive days in March 2015 and April 2015 (Porto), frozen immediately after collection to prevent degradation of the compounds and sent to Milan within 24 h in cooler boxes with dry ice or ice packs to keep them frozen. Samples were stored at −20 °C until analysis. For each sample the flow rate of the sewage stream (L/day) was recorded.

2.3. Extraction and analysis

Before solid phase extraction, samples were thawed in a warm bath, then filtered to remove suspended particulate matter through a 1.6 μm GF/A glass microfiber filters and 0.45 μm mixed cellulose membrane filters from Whatman (Kent, UK). Then 3 mL of filtered wastewater were spiked with labeled internal standards (20 ng of caffeine-13C3 and 200 ng 1,7-dimethyluric acid d6) and, if necessary, the pH was adjusted to 6.0–7.5 with 12% HCl (v/v). Samples were loaded on Oasis HLB cartridges (3 mL, 60 mg), previously conditioned with 6 mL of MeOH and 3 mL of water. Cartridges were vacuum-dried for 10 min, wrapped in aluminum foil and immediately stored at −20 °C. For analysis, cartridges were eluted with 2 mL of methanol and the extract was evaporated to dryness under a nitrogen stream. Dry residues were redissolved in 100 μL MeOH-ultrapure water (20:80, v/v), centrifuged and transferred into glass vials for instrumental analysis. One μL of the final extract was injected into the liquid chromatography coupled to tandem mass spectrometry system (LC–MS/MS). The analyses were done by high-performance liquid chromatography (1200 Series pumps system, Agilent Technologies, CA) coupled to a triple quadrupole mass spectrometer (AB SCIEX QqQ 5500, Ontario, Canada). Samples were analyzed using the positive electrospray ionization mode. Experimental conditions and detailed analytical conditions are described in Table S3 and S4 and in more detail in Sentia et al. (2015).

2.4. Daily mass loads and back-calculation of consumption

The daily mass loads (g/day) of the selected analytes were calculated multiplying the measured concentrations of caffeine and metabolites (ng/L) by the daily flow rate of wastewater (L/day) at the entry of each WWTP.

Caffeine consumption was back-calculated using the approach proposed for illicit drugs by Zuccato et al., 2008. Specific correction factors were developed taking into account the percentage of urinary excretion of each metabolite and the molar mass ratio of the parent compound to the metabolite. All the pharmacokinetic studies accessible in the literature which reported data on the human urinary excretion of caffeine after oral administration (eight in all, see SI) were reviewed to develop a specific correction factor for back-calculating caffeine intake by the population. The mean percentage of excretion of caffeine and its metabolites was calculated by weighting the number of subjects in each study. The total uncertainty related to the back-calculation procedure was evaluated as the standard deviation (SD) of the mean percentage of excretion (Table 1). This method had been previously proposed for refining the correction factors of the most used illicit drugs (Castiglioni et al., 2013; Gracia-Lor et al., 2016).

3. Results and discussion

3.1. Caffeine biomarkers for back-calculation

Selecting a substance as a biomarker is not easy to achieve as it must have specific characteristics (Gracia-Lor et al., 2016): i) be excreted in measurable quantities in wastewater; ii) be released to sewers exclusively from human excretion; iii) be unique to human metabolism to ensure that it comes only from human excretion and not from exogenous sources; iv) have low adsorption for suspended particulate; v) be stable in wastewater during in-sewer transport, and during storage and analysis.

Each substance for this investigation was tested as a suitable biomarker of caffeine consumption as described above. Caffeine itself is not a good candidate because it comes not only from coffee but also from other sources. Caffeine metabolites too may originate from other naturally occurring alkaloids with similar structures, such as theobromine and theophylline, which themselves are also caffeine metabolites (Fig. 1). Theobromine is present in cocoa beans (and subsequently in chocolate), tea leaves and cola beans. Theophylline is present in tea leaves in small amounts but is also used medically, for instance for asthma and other lung diseases (Senchina et al., 2014). Specifically, among five caffeine metabolites studied, 1-methylxanthine and 1-methyluric acid are also metabolites of theophylline, while 7-methylxanthine is the major metabolite of theobromine. Paraxanthine and 1,7-dimethyluric acid however, are exclusively metabolites of caffeine (Fig. 1). Thus, they

<p>| Table 1 |
|--------------------|------------------|------------------|------------------|</p>
<table>
<thead>
<tr>
<th>Compound</th>
<th>Mean excretion (%) from pharmacokinetic studies (SD)</th>
<th>Geometric mean from spot urine analysis (95%CI) (2466 subjects)</th>
<th>Mean excretion (%) from wastewater analysis (SD) (70 samples)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Caffeine (1,3,7-trimethylxanthine)</td>
<td>1.7 (1.0)</td>
<td>1.81 (1.57–2.08)</td>
<td>20.9 (6.0)</td>
</tr>
<tr>
<td>Paraxanthine (1,7-dimethylxanthine)</td>
<td>4.6 (1.4)</td>
<td>7.47 (6.73–8.29)</td>
<td>22.1 (4.0)</td>
</tr>
<tr>
<td>1-Methylxanthine</td>
<td>10.0 (3.4)</td>
<td>17.1 (15.4–19.0)</td>
<td>15.8 (3.5)</td>
</tr>
<tr>
<td>7-Methylxanthine</td>
<td>3.1 (1.2)</td>
<td>31.4 (28.6–34.3)</td>
<td>24.9 (6.4)</td>
</tr>
<tr>
<td>1-Methyluric acid</td>
<td>16.5 (6.2)</td>
<td>39.4 (35.8–43.4)</td>
<td>47.1 (1.1)</td>
</tr>
<tr>
<td>1,7-Dimethyluric acid</td>
<td>6.7 (2.3)</td>
<td>12.2 (11.0–13.6)</td>
<td>11.6 (2.0)</td>
</tr>
<tr>
<td>Theophylline (1,3-dimethylxanthine)</td>
<td>0.6 (0.4)</td>
<td>0.872 (0.796–0.955)</td>
<td>Not analyzed</td>
</tr>
<tr>
<td>Theobromine (3,7-dimethylxanthine)</td>
<td>1.5 (1.3)</td>
<td>12.4 (11.4–13.5)</td>
<td>Not analyzed</td>
</tr>
<tr>
<td>1,3-Dimethyluric acid</td>
<td>1.6 (0.7)</td>
<td>3.51 (3.17–3.89)</td>
<td>Not analyzed</td>
</tr>
<tr>
<td>3,7-Dimethyluric acid</td>
<td>0.2 (0.4)</td>
<td>0.784 (0.714–0.861)</td>
<td>Not analyzed</td>
</tr>
<tr>
<td>3-Methylxanthine</td>
<td>2.0 (1.1)</td>
<td>19.2 (17.5–21.0)</td>
<td>Not analyzed</td>
</tr>
</tbody>
</table>

* Data taken from Rybak et al, 2014
are potentially the most suitable biomarkers to back-calculate the amount of caffeine consumed, i.e., the consumption of all products containing caffeine (coffee, chocolate, tea, etc.). As they come only from human excretion and not from exogenous sources, their presence can play an important role in identifying fresh water or ground water contaminated by sewage.

3.2. Metabolic profiles in wastewater and in human urine

According to the human urinary excretion profile of caffeine, the mass loads of 1-methyluric acid should be the highest, followed by 1-methylxanthine, 1,7-dimethyluric acid, paraxanthine, 7-methylxanthine and finally, caffeine (Table 1). However, the quantitative profiles of caffeine and the metabolites calculated from wastewater analysis did not completely agree with the human excretion profile. The mass loads (mean of the ten cities) decreased as follows: 7-methylxanthine > paraxanthine > caffeine > 1-methylxanthine > 1,7-dimethyluric acid > 1-methyluric acid (Fig. 2). Hence, there are large differences from the human excretion profile of caffeine. We therefore included supplementary data from spot urine analysis in our comparison (Table 1). These percentages (geometric mean, 95% CI) were obtained from Rybak et al., 2014, who recently measured caffeine and 14 metabolites in >2000 urine samples. We calculated also the percentages of excretion using the concentrations measured in wastewater in the ten European cities (Table 1). Each metabolite is reported as a percentage of the sum of the levels of metabolites plus caffeine measured in wastewater, following the procedure employed by Castiglioni et al., 2011 to calculate the metabolic profile of cocaine in wastewater and in human urine. The excretion profiles of caffeine and its metabolites were calculated using median values because of the high variability of the concentrations.

Data from wastewater could be reasonably compared with the profiles in spot urine samples, since they indicate respectively the profiles of excretion from an entire community and from single individuals. Percentages were comparable for 1-methylxanthine and 7-methylxanthine acid in wastewater and spot urine samples, but higher than in pharmacokinetic studies (Table 1). This can be easily explained by the fact that

Fig. 1. Metabolic pathway of caffeine in humans.
they are also metabolites of theophylline and theobromine respectively. The percentage of caffeine in wastewater (21%) was much higher than expected from spot urine analysis and pharmacokinetic studies (1.8% and 1.7%). There might therefore be other sources of caffeine contributing to the total amount in wastewater (e.g., coffee grounds that are disposed down the sink drain, disposal of coffee that was not drunk or improper disposal of caffeine for pharmacological use). In contrast, for 1-methyluric acid the percentage in wastewater was lower than in urine and in pharmacokinetic studies. A possible explanation could be degradation of this compound in wastewater such as in-sewer, during transport or during storage. This should be verified by in-sewer experiments and additional modeling studies.

Some differences were observed for paraxanthine (22.1% of the total in wastewater, 4.6% in pharmacokinetic studies and 7.5% in spot urine samples); however for 1,7-dimethyluric acid the results were comparable (approximately 12% of the measured concentrations in wastewater and in spot urine samples, and 4.3–12.6% of the administered dose in pharmacokinetic studies (see data in SI)). Taking to account of all these considerations, 1,7-dimethyluric acid seemed to be the most suitable biomarker for the back-calculation of caffeine. The mean percentage of excretion of this metabolite weighted by the number of subjects in each study (6.7%) and the caffeine/1,7-dimethyluric acid molecular mass ratio were used to obtain the correction factor (CF), according to the following equation:

$$\text{CF} = \frac{\text{M}_w\text{caffeine}}{\text{M}_w\text{1,7-dimethyluric acid}} \times \frac{\text{Mean excretion}_{\text{1,7-dimethyluric acid}}}{0.067} = 14.8$$

where \(\text{M}_w\) is the molecular weight and the mean excretion is the weighted mean of the percentage of excretion of the target metabolite.

### 3.3. Estimation of caffeine consumption

Using the proposed correction factor, caffeine consumption (in mg/day/person) in each city was calculated based on the wastewater measurements of 1,7-dimethyluric acid. The mean daily consumption of caffeine per capita ranged from 263 mg/day/person in Zurich to 87 mg/day/person in Milan (Table 2). These data match the mean daily caffeine intake in Europe of around 300 mg/day/person estimated by the European Food Safety Authority (means range from 37 to 320 mg/day/person estimated from individual surveys for adults between 18 and 64 years) (European Food Safety Authority (EFSA), 2015).

### Table 2

Caffeine consumption estimated from wastewater analysis and using coffee trade data for the countries investigated. The difference was calculated between the estimates from international statistics and from wastewater analysis.

<table>
<thead>
<tr>
<th>Cities investigated (country)</th>
<th>Caffeine from wastewater analysis</th>
<th>Caffeine from international statisticsb</th>
<th>Difference (%)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Bristol (UK)</td>
<td>190 (37)</td>
<td>3.3</td>
<td>−38</td>
</tr>
<tr>
<td>Brussels (Belgium)</td>
<td>162 (15)</td>
<td>4.3</td>
<td>16</td>
</tr>
<tr>
<td>Castellón (Spain)</td>
<td>122 (28)</td>
<td>4.5</td>
<td>53</td>
</tr>
<tr>
<td>Copenhagen (Denmark)</td>
<td>229 (19)</td>
<td>6.9</td>
<td>−16</td>
</tr>
<tr>
<td>Lugano (Switzerland)</td>
<td>97 (16)</td>
<td>7.6</td>
<td>55</td>
</tr>
<tr>
<td>Milan (Italy)</td>
<td>86 (18)</td>
<td>5.6</td>
<td>63</td>
</tr>
<tr>
<td>Oslo (Norway)</td>
<td>211 (21)</td>
<td>8.7</td>
<td>15</td>
</tr>
<tr>
<td>Porto (Portugal)</td>
<td>121 (27)</td>
<td>4.8</td>
<td>56</td>
</tr>
<tr>
<td>Utrecht (The Netherlands)</td>
<td>107 (28)</td>
<td>5.3</td>
<td>51</td>
</tr>
<tr>
<td>Zurich (Switzerland)</td>
<td>263 (23)</td>
<td>7.6</td>
<td>−20</td>
</tr>
</tbody>
</table>

a (Garattini, 1993).

For a more accurate comparison, we compared our wastewater analysis data to the amount of coffee consumed per country per capita (per person on average), which reflects the imports of coffee by each country, according to the International Coffee Organization (ICO) (2015). We converted the per capita consumption (in kg/person) of coffee to the daily intake of caffeine per person considering that dry coffee beans contain about 1.1% of caffeine in Arabica and about 2.2% in Robusta coffee. In 2015, around 60% of the coffee exported was Arabica (International Coffee Organization, 2015), but the proportion can change from country to country. For instance, according to Garattini, 1993, consumer countries can be classified in three levels: (a) where consumption of Arabica accounts for >70% (Switzerland and Northern European countries, i.e. Norway and Denmark); (b) where consumption of Arabica is around 50% (Italy, the Netherlands, Belgium and the UK); (c) where consumption of Robusta predominates (Spain and Portugal) (Table 2). In addition, the amount of caffeine extracted varies with the preparation method, ranging from 75% in boiled coffee to nearly 100% in filtered coffee. To estimate the amount of caffeine in the coffee we took 1.1% for countries classified in group (a), 1.6% (i.e. mean caffeine content in Arabica and in Robusta) for countries belonging to group (b) and 2.2% for countries in group (c). In all cases, we assumed 95% extraction efficiency, as previously proposed (Fredholm et al., 1999).

For four cities (Oslo, Copenhagen, Zurich and Brussels), the difference was 20% or less. The amounts for Castellón, Utrecht, Milan, Lugano and Porto estimated from wastewater analysis were lower than indicated by the coffee trade figures, and higher in Bristol. This might be due to different factors: first of all, we compared data from whole country with data in a specific city, while population habits might be different. This was the case for Zurich and Lugano, two Swiss cities: a 20% difference was obtained for Zurich (410,000 inhabitants) whilst it was around 50% for Lugano (100,000 inhabitants). Secondly, we compared annual coffee trade figures with caffeine estimated through wastewater analysis in one week. Finally, data obtained through back-calculation refer to the amount of caffeine consumed in all products that contain relatively large amounts such as coffee, chocolate, soft drinks and medications. Thus, larger amounts of caffeine estimated through the wastewater analysis in Zurich, Copenhagen, and especially in Bristol, might be due to higher consumption of other products in those countries. Switzerland is in fact the country with the highest per capita consumption of chocolate, and the UK is also among the countries with the highest consumption, according to different sources (Statista, 2015; Target Map, 2015). Another reason might be the fact that the caffeine content of coffee in the UK is higher than in other countries (Barone and Roberts, 1996). Furthermore, tea containing around 3% of caffeine is the most popular drink in the UK today, and contributes to caffeine consumption. In five cities, the difference was of at least 50%.

The aim of the comparison between the amount of caffeine consumed, estimated from the wastewater analysis, and coffee consumption figures from international trade was mainly to check whether the proposed metabolite was a suitable biomarker of consumption. The results indicate that 1,7-dimethyluric acid can be used for this purpose, although additional studies are needed to validate this approach, including more extensive wastewater sampling campaigns in different countries.

Additional information on the current proportions (percentages) of commercial varieties of coffee consumed in each country is also needed for more accurate comparisons. There are some differences between coffee consumption data, in terms of the amount consumed in each country per capita, published by different sources (for instance, between the ICO (International Coffee Organization (ICO), 2015) which is based on coffee imports and exports and Euromonitor International (Caffeine Informer, 2016), which deals with local business information). This is another factor that may influence the accuracy of a data comparison.

Additionally, only eight studies could be found dealing with the human excretion of caffeine, so more pharmacological studies are essential to improve the reliability of urinary excretion profiles and the correction factors used to back-calculate caffeine consumption. At present, these studies are scarce and most are quite old and based on a small number of subjects (Gracia-Lor et al., 2016).

4. Conclusions

Profiles of caffeine metabolites in wastewater reasonably matched the profiles in spot urine samples suggesting that the analysis in wastewater might reflect the collective consumption of caffeine-containing products.

We selected 1,7-dimethyluric acid for caffeine back-calculation because it is an exclusive human metabolite of caffeine and so it is only produced by consumption of products containing caffeine (i.e. coffee, tea, chocolate, etc.). The percentage of its excretion from pharmacokinetic studies is similar to the profiles found in urine and in wastewater (estimated from 70 influent wastewater samples collected in ten European cities). The mean daily consumption of caffeine per capita, estimated from wastewater analysis using the correction factor proposed, matched the mean daily caffeine intake (from 37 to 320 mg/day/person estimated from individual surveys for adults 18–64 years old). In four cities a good correlation was seen between wastewater analysis and the amount of coffee consumed in the country per capita. Several factors might explain discrepancies in the other six cities. For instance the estimation of coffee consumption on the basis of the imports of coffee by each country is influenced by many uncertainties, so it is hard to estimate the consumption of other commodities contributing to caffeine intake. Furthermore, the correction factor may be imprecise due to uncertainties in the metabolism studies in the literature. Thus, new studies are needed about the metabolism and urinary excretion of caffeine in realistic intake amounts. Stability tests of biomarkers in sewers are also needed.

Contributions

Emma Gracia-Lor, Ettore Zuccato and Sara Castiglioni planned and designed the study. The collection of the wastewater samples was organized by all authors. Emma Gracia-Lor analyzed the samples and interpreted the results with the input of Nikolaos I. Rousis and Sara Castiglioni. Emma Gracia-Lor drafted the manuscript, which was critically revised by all co-authors. All authors are aware of the content, and accept responsibility, for the manuscript.

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Appendix A. Supplementary data

Supplementary data to this article can be found online at http://dx.doi.org/10.1016/j.scitotenv.2017.07.258.