Reef sponges facilitate the transfer of coral-derived organic matter to their associated fauna via the sponge loop


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INTRODUCTION

Highly biodiverse tropical coral reefs are characterized by complex trophic webs and intricate networks of positive and negative interspecific interactions that are fundamental for understanding reef ecosystem functioning (Glynn 2004, Briand et al. 2016, Harborne et al. 2017). Due to high levels of predation and intense competition for space and resources, the exploitation of other organisms as living habitat is a common strategy on coral reefs (e.g. Duffy 1992, Munday et al. 1997, Elliott & Mariscal 2001). Sponges, an abundant faunal component of coral reefs, are amongst the most widely exploited, harbouring an exceptionally diverse array of crustaceans, molluscs, bryozoans, polychaetes, cnidari-

ABSTRACT: The high biodiversity of coral reefs results in complex trophic webs where energy and nutrients are transferred between species through a multitude of pathways. Here, we hypothesize that reef sponges convert the dissolved organic matter released by benthic primary producers (e.g. corals) into particulate detritus that is transferred to sponge-associated detritivores via the sponge loop pathway. To test this hypothesis, we conducted stable isotope (13C and 15N) tracer experiments to investigate the uptake and transfer of coral-derived organic matter from the sponges *Mycale fistulifera* and *Negombata magnifica* to 2 types of detritivores commonly associated with sponges: ophiuroids (*Ophiothrix savignyi* and *Ophiocoma scolopendrina*) and polychaetes (*Polydorella smurovi*). Findings revealed that the organic matter naturally released by the corals was indeed readily assimilated by both sponges and rapidly released again as sponge detritus. This detritus was subsequently consumed by the detritivores, demonstrating transfer of coral-derived organic matter from sponges to their associated fauna and confirming all steps of the sponge loop. Thus, sponges provide a trophic link between corals and higher trophic levels, thereby acting as key players within reef food webs.

KEY WORDS: Coral mucus · Reef trophic web · Detritus · Sponge loop · Detritivore · Trophic interactions · Interspecific associations
ans, echinoderms (including ophiuroids), and fishes (e.g. Pearse 1950, Westinga & Hoetjes 1981, Duarte & Nalessso 1996, Wulff 2006). Many of these associations are opportunistic and transient, but others are obligate, or even species-specific (Henkel & Pawlik 2005), forming relationships ranging from mutualism (Meroz & Ilan 1995) to parasitism (Pawlik 1983, Đuriš et al. 2011). However, most of these relationships, including potential trophic interactions, are poorly characterized.

Sponges not only offer habitat and physical protection for their associates, but can also provide food through predation on sponge tissue (Pawlik 1983, Oshel & Steele 1985, Đuriš et al. 2011), exploitation of the enhanced particle flow induced by sponge pumping (Westinga & Hoetjes 1981, Costello & Myers 1987), or deposit-feeding on detritus that settles on the sponge surface (Hendler 1984, Henkel & Pawlik 2005). Sponges can also actively generate detritus that may be utilized by reef fauna (Hammond & Wilkinson 1983, de Goeij et al. 2013). Recently, it was hypothesized that by taking up dissolved organic matter (DOM) and converting it into particulate detritus, sponges enable the transfer of the energy and nutrients in DOM to higher trophic levels on coral reefs via a pathway defined as the ‘sponge loop’ (de Goeij et al. 2013, Rix et al. 2016, 2017). Benthic primary producers are the main producers of labile DOM on reefs, as they release large quantities of the carbon they fix into the surrounding water as DOM (Crossland 1987, Barrón & Duarte 2009, Haas et al. 2011). Corals, for example, devote up to ~40% of their net photosynthetic output into the release of coral mucus (Crossland et al. 1980, Haas et al. 2011, Tremblay et al. 2012). This mucus is released in both dissolved and particulate forms (Crossland 1987, Naumann et al. 2010), although the majority of the particulate mucus also subsequently dissolves into the surrounding water, further contributing to the reef DOM pool (Wild et al. 2004). This DOM pool represents one of the largest available organic matter pools on corals reefs, and yet its energy and nutrients remain largely inaccessible to most reef fauna as a major food source. Reef sponges, however, are not only able to take up natural reef DOM at high rates, but DOM can account for the majority (up to ~90%) of their total heterotrophic carbon uptake (Yahel et al. 2003, de Goeij et al. 2008, Mueller et al. 2014a, McMurray et al. 2016, Morganti et al. 2017). Furthermore, up to ~40% of the DOM assimilated by sponges is subsequently released as detritus (de Goeij et al. 2013, Rix et al. 2016, 2017), a substrate that is consumed by a wide range of reef fauna (e.g. Glynn 2004). Sponges therefore convert reef DOM into a food source that would be more readily accessible to their detritus-feeding associated fauna. Consequently, we hypothesize that sponges not only generate food for their associated detritivores through the production of sponge detritus, but also provide a direct trophic link between corals and sponge-associated detritivores that allows these associates ready access to the dissolved energy and nutrients produced by corals.

To test this hypothesis, we conducted 2 stable isotope tracer experiments with $^{13}$C- and $^{15}$N-labelled corals to follow the transfer of coral-derived organic matter (i.e. coral mucus) through each step of the sponge loop: (1) uptake and assimilation of naturally released coral mucus by the sponges, (2) release of assimilated coral mucus by the sponges as sponge detritus and (3) uptake of sponge detritus by 2 types of detritivores commonly associated with reef sponges: polychaetes and ophiuroids. First, we investigated the trophic transfer of coral mucus through the branching sponge Negombata magnifica to its associated spionid polychaete Polydorella smurovi. Spionid polychaetes are deposit and suspension feeders lacking mechanisms for predatory feeding, and most known species of Polydorella are associated with sponges (Taghon et al. 1980, Dauer et al. 1981, Williams & McDermott 1997, Williams 2004). Secondly, we examined the transfer of coral mucus through the encrusting sponge Mycale (Carmia) fistulifera to the detritus-feeding ophiuroids Ophiothrix savignyi and Ophiocoma scolopendrina (Magnus 1965, Warner & Woodley 1975, Hendler 1984), which commonly inhabit sponges (O. savignyi) (James & Pearse 1969) and ruble on reef flats (O. scolopendrina) (Magnus 1965).

**MATERIALS AND METHODS**

**Study site and organism collection**

This study was conducted at the Marine Science Station (MSS) in Aqaba, Jordan (northern Gulf of Aqaba, Red Sea; 29°27’N, 34°58’E) during September and October 2013. Sampling was carried out on the ~1 km long fringing reef in front of the MSS between 8 and 20 m water depth by SCUBA. Free-living fungiid corals (genera: Fungia, Ctenactis and Herpolitha; n = 30) were collected as they can be removed from the reef without physical damage and produce large quantities of coral mucus (Naumann et al. 2010). Corals were transferred to the MSS without air exposure and maintained in running-seawater facilities
(>1000 l) for at least 72 h prior to the start of experiments. Two abundant sponge species were collected, the encrusting sponge *Mycella (Carmia) fistulifera* (20 ± 8 cm² fragments were chiselled from dead branching corals) and the branching sponge *Negombata magnifica* (small branches of 67 ± 22 cm² were cut), and maintained in 100 l flow-through aquaria for at least 1 wk of acclimation. *Polydorella smurovi* polychaetes were collected by cutting branches of densely infested *N. magnifica* specimens. The ophiuroids were collected from sponges (*Ophiolithrix savignyi*) and the reef flat (*Ophiocoma scolopendrina*). *O. scolopendrina* was included for comparison of a typically non-sponge-associated detritivore that utilizes a similar feeding mode (Magnus 1965, and therefore may be representative of a strictly transient or opportunistic associate that only rarely encounters sponges). Detritivores were acclimated with their host sponges for 48 h prior to experimentation.

**Stable isotope tracer experiment (coral–coral mucus–sponge–detritus–detritivore)**

Stable isotope (¹³C and ¹⁵N) labelling of corals was conducted over 8 d as described by Rix et al. (2016). Briefly, corals were labelled in 100 l aquaria to which the seawater flow-through was stopped and 36 mg l⁻¹ NaH¹³CO₃ and 1 mg l⁻¹ Na¹⁵NO₃ (Cambridge Isotopes, 98% ¹³C and ¹⁵N) were added. Aquarium pumps maintained water circulation and gas exchange until seawater flow-through was resumed overnight (8 h). Water temperature was maintained within ±1°C of *in situ* conditions by placing aquaria in a flow-through raceway (flow rate ~1000 l h⁻¹). After the final day of labelling, the corals were transferred to the raceway and rinsed in fresh flowing seawater overnight to remove any unincorporated label. Coral mucus was collected from a sub-set of the corals (n = 6) by brief (2 min) air exposure and frozen at −80°C for subsequent stable isotope analysis. Corals were transferred to the experimental set-up for the subsequent tracer experiment the following morning.

The trophic transfer of mucus from the ¹³C- and ¹⁵N-labelled corals through the sponge tissue and detritus to sponge-associated detritivores was investigated using six 2-tiered, flow-through aquaria setups, consisting of paired upper and lower aquaria (100 l each) connected via constant water flow. The upper aquaria (light levels ~120 µmol quanta m⁻² s⁻¹) were supplied with fresh-pumped reef water at a rate of ~10 l min⁻¹, which then flowed into the lower aquaria. Labelled corals (n = 10 per aquarium) were maintained in 3 of the upper aquaria, while the remaining 3 upper aquaria served as coral-free controls. The lower aquaria contained sponges cleared of their associated detritivores; either *N. magnifica* or *M. fistulifera* (n = 4 per aquarium). The set-up was designed to mimic natural *in situ* conditions as closely as possible, and corals were allowed to release mucus at natural rates without manipulation. Thus, here we consider the transfer of bulk mucus and do not differentiate between particulate or dissolved fractions. To ensure conditions for the first step of the sponge loop were met (i.e. the uptake of dissolved mucus by the sponges), samples for dissolved organic carbon (DOC) and bacterioplankton were taken from the upper aquaria (n = 3–6 per aquarium and n = 9–18 per treatment) to compare differences between the treatment and control aquaria to verify DOC release by the corals. To determine DOC uptake by the sponges, the flow-through from the coral aquaria to the sponge aquaria was briefly stopped (30 min) and initial DOC and bacterioplankton samples were taken from each sponge tank and then resampled after 30 min to measure uptake by the sponges (n = 3 per aquaria, n = 9 per treatment). The flow rate to the upper aquaria (~10 l min⁻¹) ensured the set-up was replaced with fresh seawater every 10 min in order to supply the sponges with sufficient food as well as to prevent bacterioplankton growth and potential bacterial-mediated transfer of coral mucus to the sponges. After 5 d exposure to seawater flowing from the aquaria containing the labelled corals, 1 sponge per tank was removed, rinsed in label-free seawater for 10 min, and frozen at −80°C for stable isotope analysis (n = 3 per treatment). The corals were removed and all aquaria were thoroughly cleaned and flushed with fresh flowing seawater for 2 h to eliminate any labelled organic matter originating from the corals prior to introducing the detritivores in order to ensure any subsequent enrichment of ¹³C and ¹⁵N could be attributed to the sponges. Detritivores were then transferred onto the remaining experimental sponges. *P. smurovi* specimens were transferred with a pipette onto the surface of *N. magnifica* where they quickly re-established themselves on the sponge surface. The ophiuroids *O. savignyi* and *O. scolopendrina* (n = 4 per aquarium, n = 12 per treatment) were introduced to the aquaria with *M. fistulifera* and immediately took refuge in crevices inside the sponges. One and 5 d after the addition of the detritivores, detritus was collected from the surface of each sponge with a pipette, pooled by aquarium for stable isotope analysis (n = 2 per aquarium, n = 6 per treat-
ment), and frozen at −80°C for isotope analysis. After 5 d, the polychaetes, ophiuroids, and remaining sponges were frozen at −80°C for isotope analysis. Due to their small size, polychaetes from each aquarium were pooled onto 1 pre-combusted (450°C, 4 h) GF/F filter (n = 3 per treatment).

**Sponge detritus production**

To determine detritus production, *N. magnifica* and *M. fistulifera* specimens were incubated individually in stirred 2 l chambers for 3 h (n = 6 per species) as previously described (Rix et al. 2016). To prevent contamination with previously accumulated detritus or sediment, the sponges were carefully cleaned of all detritus and debris using gentle suction with a small tube a few mm from the sponge surface without touching or disrupting the sponges. Sponges were then transferred without air exposure to the base plate of the incubation chambers and cleaned again prior to closing the chambers. Incubations without sponges (n = 6) served as controls. Initial samples for particulate organic carbon (POC) and nitrogen (PON) were taken at the start of the incubation from the fresh seawater used to fill the chambers (n = 6). At the end of the incubation, sponges were carefully removed, and the incubation water was homogenized and 1 l gently vacuum-filtered onto 2 separate pre-combusted GF/F filters (1 each for POC and PON). Filters were dried at 40°C for at least 48 h and stored dry until C:N elemental analysis. Sponge surface area and thickness were measured to determine the sponge volume. Fluxes of POC and PON were corrected for the initial concentrations of POC and PON in the seawater, and the sponge fluxes were corrected for differences with seawater controls. Rates were then normalized to sponge volume and incubation time and presented as µmol C or N cm⁻³ sponge d⁻¹.

**DOC and flow cytometry measurements**

DOC samples were collected in pre-cleaned 60 ml syringes and gently vacuum filtered (maximum pressure 20 kPa) through pre-combusted GF/F filters directly into 30 ml high-density polyethylene (HDPE) sample bottles using a customized set-up. Syringes, filtration apparatus, and sample bottles were acid-washed in 0.4 M HCl for 24 h and rinsed twice with Milli-Q water before sampling. The first 20 ml of sample water was used to rinse the sample bottles (2 × 10 ml). The remaining 30 ml was collected, acidified with 80 µl of 18.5% HCl, and stored at 4°C in the dark until analysis. Samples were measured using high-temperature catalytic oxidation (HTCO) on a total organic carbon analyser (Shimadzu TOC-VCPH). The instrument was calibrated with a 10 point calibration using serial dilutions of potassium hydrogen phthalate (certified stock solution 1000 ppm Standard Fluka 76067). Deep sea reference (DSR) water standards (Batch 13, 41–45 µmol C l⁻¹) supplied by the Consensus Reference Material (CRM) Project (Hansell Lab, University of Miami) were applied as a positive control after every 10 samples to determine the accuracy and precision of the instrument. Each sample was averaged over 5 measurements and analytical precision was <3% of the certified value. Bacterioplankton samples (2 ml) were fixed in 0.1% paraformaldehyde (final concentration) for 30 min at room temperature, frozen with liquid N, and stored at −80°C until analysis. Abundances of heterotrophic bacteria were quantified on a FACSCalibur flow cytometer (Becton Dickinson, 488 nm excitation laser). Samples were stained with SYBR Green 1 for 30 min prior to sorting at a flowrate of approximately 0.06 µl min⁻¹ for 1 min. Heterotrophic bacteria were gated on a plot of side scatter versus green fluorescence using CellQuestPro (BD Biosciences). The cytometer flow rate was gravimetrically calibrated according to Marie et al. (1999), and all samples were measured on the same day.

**Sample treatment and stable isotope analysis**

Sponge tissue, sponge detritus, and ophiuroid tissue samples were lyophilized and homogenized, and subsamples were weighed into silver cups for δ¹³C and δ¹⁵N analysis. Samples for δ¹³C were decalcified with 0.4 M HCl to obtain the organic carbon (Corg) fraction. GF/F filters (polychaetes and filters from sponge detritus production incubations) were decalcified in an atmosphere of fuming HCL, re-dried at 40°C and folded into silver cups. Isotope ratios and C:N content were measured simultaneously using a Flash 1112 EA coupled to a Delta V IRMS via a ConFlo IV- interface (Thermo Scientific). Standard deviations of C and N content were <3% of the concentrations analysed and <0.15‰ for repeated δ¹³C and δ¹⁵N measurements of standard material (peptone). Carbon and nitrogen stable isotope ratios are expressed in delta notation as: δ¹³C or δ¹⁵N (‰) = (Rsample / Rref − 1) × 1000, where R is the ratio of ¹³C:¹²C or ¹⁵N:¹⁴N in the sample or reference mater-
rial: Vienna Pee Dee Belemnite for C (\(R_{\text{ref}}=0.01118\)) and atmospheric nitrogen for N (\(R_{\text{ref}}=0.00368\) N). The incorporation of excess (i.e. above background) \(^{13}\)C and \(^{15}\)N tracer (see Fig. 1) was calculated by subtracting the background \(\delta^{13}\)C and \(\delta^{15}\)N values of the control samples from the treatment samples as follows: \(\Delta \delta^{13}\)C = \(\delta^{13}\)C\text{sample} - \(\delta^{13}\)C\text{background}\) and \(\Delta \delta^{15}\)N = \(\delta^{15}\)N\text{sample} - \(\delta^{15}\)N\text{background}\). In order to calculate the uptake rates of coral mucus by the sponges and sponge detritus by the detritivores (see Fig. 2), the excess fractional abundance of heavy isotope (\(E\)) in the sponge and detritivore tissue was calculated as: \(E = F_{\text{sample}} - F_{\text{background}}\) where \(F_{\text{sample or background}} = \frac{R_{\text{sample or background}}}{(R_{\text{sample or background}} + 1)}\). The total uptake (\(I\)) of \(^{13}\)C and \(^{15}\)N was then calculated by multiplying \(E\) by the C\(_{\text{org}}\) or N content of the sponge or detritivore tissue. To determine the total C (\(^{12}\)C + \(^{13}\)C) and N (\(^{14}\)N + \(^{15}\)N) incorporated, \(I\) was divided by the fractional abundance of either the coral mucus (to determine total uptake rates of coral mucus C and N into the sponge tissue) or detritus (to determine the total uptake rates of sponge detritus C and N by the detritivores). Rates were then normalized to time and tissue C\(_{\text{org}}\) or N content of the sponges or detritivores (see Fig. 2).

**Data analysis**

Statistical analyses to determine differences in \(^{13}\)C and \(^{14}\)N enrichment in the control and treatment samples were conducted in PRIMER-E version 6 (Clarke & Gorley 2006) with the PERMANOVA+ add-on (Anderson et al. 2008) using individual 1-factor permutational multivariate analyses of variance (PERMANOVAs) with Type III (partial) sum of squares and unrestricted permutation of raw data (999 permutations). PERMANOVAs with Monte Carlo tests were used when the sample size could not provide sufficient permutations (i.e. \(n = 3\) for polychaetes, Anderson et al. 2008). Generalised linear models (GLM) were used to compare concentrations of DOC and bacterioplankton in the treatment and control aquaria. First, we tested for a potential tank effect within each treatment by running a GLM with ‘Tank’ as the fixed factor. Since this was not significant, an additional GLM was run with ‘Treatment’ as the single fixed factor. Assumptions of normally distributed and homogenous residuals were confirmed using QQ plots and scatter plots of residuals against fitted values, and data were transformed where necessary. Paired \(t\)-tests were used to determine differences in DOC and bacterioplankton concentrations in the sponge aquaria before and after the flow-through seawater was stopped for 30 min. The treatment and control aquaria were tested separately as they showed different trends for DOC. These statistical tests were carried out in R v. 3.3.3 (R Core Team 2012).

**RESULTS**

DOC concentrations were significantly elevated in the treatment aquaria containing the labelled corals (mean ± SD, 83.7 ± 4.9 µM) compared to the control aquaria without corals (76.6 ± 4.7 µM), demonstrating release of DOC by the labelled corals (GLM: \(F_{1,52} = 29.1, p < 0.001\)). By contrast, there were no significant differences in the concentrations of bacterioplankton in the treatment and control aquaria (1.84 ± 0.29 × 10\(^5\) and 1.83 ± 0.20 ×10\(^5\) cells ml\(^{-1}\), respectively; \(F_{1,34} = 0.05, p = 0.8203\)), indicating the flow rate was sufficient to prevent coral mucus-fuelled growth of bacteria in the aquaria. When the flow-through seawater to the sponge aquaria was stopped, there was a significant decrease in bacterioplankton concentrations in both the treatment (−0.04 ± 1.22 × 10\(^5\) cells ml\(^{-1}\)) and control (−4.59 ± 2.10 × 10\(^4\) cells ml\(^{-1}\)) aquaria after 30 min, demonstrating the sponges were actively filtering (paired \(t\)-test: \(t = 16.7, df = 8, p < 0.001\) and \(t = 8.4, df = 8, p < 0.001\) for the treatment and control aquaria, respectively). However, DOC concentrations showed a significant decrease only in the treatment aquaria after 30 min (net DOC flux: −6.5 ± 4.4 µM DOC; paired \(t\)-test: \(t = 5.63, df = 17, p < 0.001\)), while no changes were observed in the control aquaria (net DOC flux: +1.44 ± 5.5 µM; \(t = −1.11, df = 17, p = 0.2831\)). DOC removal by the treatment sponges (6.5 ± 4.4 µM) corresponded with the increase in DOC concentration between the treatment and control aquaria (7.1 µM). Thus, the consistent net uptake of DOC only by sponges in the treatment aquaria, where DOC concentrations were initially elevated due to release of DOC by the labelled corals, demonstrates uptake of coral-derived DOC.

After labelling with the stable isotope tracers (\(\text{NaH}_{13}\)CO\(_3\) and \(\text{Na}_{15}\)NO\(_3\)), the corals produced mucus that was enriched in both \(^{13}\)C and \(^{15}\)N (Fig. 1). The stable isotope tracer experiments confirmed the transfer of this coral-derived C and N into the tissue of the 2 sponges Negombata magnifica and Mycale fistulifera, as evidenced by positive (i.e. above background) \(\Delta^{13}\)C and \(\Delta^{15}\)N values after 5 d exposure to the \(^{13}\)C- and \(^{15}\)N-labelled corals (Fig. 1). Incorporation rates of coral mucus into sponge tissue were
from the experimental set-up, enrichment in $^{13}$C and $^{15}$N was detected in the detritus produced by *N. magnifica* and *M. fistulifera* (Fig. 1). Finally, after 5 d exposure to the labelled sponge detritus, enrichment of $^{13}$C and $^{15}$N was detected in the tissues of the detritivores: the polychaete *Polydorella smurovi* and the ophiuroids *Ophiothrix savignyi* and *Ophiocoma scolopendrina* (Fig. 1). In all cases, there was higher enrichment of $^{15}$N than $^{13}$C, likely due to the higher initial $^{15}$N enrichment in the coral mucus. Despite expected isotope label dilution with each trophic transfer step (e.g. due to respiration and incomplete processing of each labelled source within the time frame of the experiment), the polychaetes and ophiuroids were significantly enriched in $^{13}$C and $^{15}$N compared to the controls (polychaetes: Monte Carlo $F_{1,4} = 12.33$, $p = 0.02$ for C and $F_{1,4} = 20.62$, $p = 0.006$ for N; ophiuroids: $F_{1,24} = 15.50$, $p = 0.001$ for C and $F_{1,24} = 38.94$, $p < 0.001$ for N). Thus, coral-derived $^{13}$C and $^{15}$N were (1) released by the corals, (2) taken up by the sponges, (3) released as sponge detritus, and (4) incorporated by the detritivores. The polychaetes assimilated sponge detritus at higher rates of $32.3 \pm 13.0 \mu$mol C detritus mmol C detritivore $^{-1}$ d$^{-1}$ and $24.4 \pm 11.3 \mu$mol N detritus mmol N detritivore $^{-1}$ d$^{-1}$ compared to the ophiuroids ($7.6 \pm 6.5 \mu$mol C detritivore $^{-1}$ d$^{-1}$ and $6.8 \pm 4.1 \mu$mol N detritivore $^{-1}$ d$^{-1}$; Fig. 2), although the SDs were high. There were no differences in rates between the 2 ophiuroid species, therefore the results were pooled.

Incubations with sponges yielded significantly higher amounts of particulate organic matter (POM) than seawater controls ($F_{1,15} = 86.02$, $p < 0.001$ and $F_{1,15} = 81.94$, $p < 0.001$ for C and N, respectively). On average, detritus production by *N. magnifica* ($15.5 \pm 7.2 \mu$mol C$_{org}$ cm$^{-3}$ d$^{-1}$ and $2.1 \pm 0.9 \mu$mol N$_{org}$ cm$^{-3}$ d$^{-1}$) was comparable to that of *M. fistulifera* ($17.2 \pm 7.5 \mu$mol C$_{org}$ cm$^{-3}$ d$^{-1}$ and $1.8 \pm 0.4 \mu$mol N$_{org}$ cm$^{-3}$ d$^{-1}$; Fig. 2). The mean (±SD) C:N ratio of the sponge detritus ($7.2 \pm 1.7$ and $6.7 \pm 1.0$ for *N. magnifica* and *M. fistulifera*, respectively) was significantly lower than that of the ambient suspended POM in the water column ($10.3 \pm 1.4$; $F_{1,16} = 17.16$, $p = 0.003$).

**DISCUSSION**

Here we show that reef sponges facilitate the transfer of coral-derived organic matter to their associated detritivores via the production of sponge detritus, thereby demonstrating all steps of the sponge loop (Fig. 3). Several sponge species are able to convert coral-derived DOM into sponge detritus (Rix et al.
likely makes up the largest fraction, as the majority (56 to 80%) of coral mucus dissolves in the water column (Wild et al. 2004), and coral loss of fixed carbon due to expulsion of Symbiodinium is typically negligible (0.01%; Hoegh-Guldberg et al. 1987) compared with mucus release (up to ~40%; Crossland et al. 1980, Tremblay et al. 2012). Coral-derived organic matter could also be indirectly transferred to sponges via bacteria, which can also consume coral mucus (Perrier-Pagès et al. 2000, Wild et al. 2010, Tanaka et al. 2011). Here, DOC measurements confirmed the uptake of coral-derived DOM by the experimental sponges, consistent with previous studies showing the dissolved fraction of coral mucus is readily taken up by several reef sponges (Rix et al. 2017). However, potential uptake of coral-derived POM or microbial-mediated transfer of coral mucus could additionally contribute to sponge recycling of coral-derived organic matter and further facilitate energy and nutrient retention within coral reefs (de Goeij et al. 2013, Rix et al. 2016). At our study site in the northern Red Sea, corals are the dominant primary producers supplying DOM to the sponge loop (Cardini et al. 2016, van Hoytema et al. 2016), but on coral reefs dominated by macroalgae, the main source of DOM for the sponge loop may rather be supplied by algae (Fig. 3). Algae not only typically release more labile DOM than corals (Haas et al. 2011, Mueller et al. 2014b), but sponges also appear to process algal-derived DOM at a higher rate than coral-derived DOM (Rix et al. 2017), suggesting increased algal cover may enhance DOM cycling though the sponge loop. Since inorganic nutrient release by sponges can additionally enhance algal growth (Slattery et al. 2013, Easson et al. 2014), it has also been hypothesized that this reciprocal coral–algae nutrient recycling may result in a positive feedback loop further promoting the growth of sponges and algae, potentially at the expense of corals (Pawlik et al. 2016). Indeed, numerous studies have highlighted how DOM released by corals versus algae exerts differing effects on reef functioning by altering microbial activity (Barott & Rohwer 2012, Haas et al. 2013, 2016, Nelson et al. 2013); however, the potential effects due to altered nutrient cycling by sponges remains poorly explored.

The encrusting sponge Mycale fistulifera was previously shown to convert coral mucus into detritus (Rix et al. 2016, 2017), but interestingly, we also found transfer of coral-derived organic matter into the detritus released by the massive branching sponge Negombata magnifica. The transfer of DOM into sponge detritus has so far only been documented
for encrusting sponges (de Goeij et al. 2013, Rix et al. 2016, 2017), and it has been suggested that high cell shedding and detritus production may be restricted to these thin encrusting species whose growth is limited by high competition for free substrate (e.g. Buss & Jackson 1979). Since the upward growth of massive sponges is not similarly constrained, they are expected to invest more resources into growth rather than high biomass turnover (Pawlik et al. 2016). Despite the differences in growth form, we found comparable detritus production rates by the 2 sponges, although with considerable intraspecific variability. Sponge detritus is hypothesized to be largely due to high cell turnover and shedding, particularly of sponge choanocyte cells (de Goeij et al. 2009, Alexander et al. 2014). However, sponges may also release detritus by ejecting waste products and incompletely digested food (Maldonado 2016) or via other mechanisms such as mucus production in response to sedimentation (Bell et al. 2015, Biggerstaff et al. 2017). Sponge cell turnover and shedding is reduced under suboptimal food conditions (Alexander et al. 2015b) and in wounded sponges (Alexander et al. 2015a), suggesting a complex interplay of factors such as food availability, predation, for encrusting sponges (de Goeij et al. 2013, Rix et al. 2016, 2017), and it has been suggested that high cell shedding and detritus production may be restricted to these thin encrusting species whose growth is limited by high competition for free substrate (e.g. Buss & Jackson 1979). Since the upward growth of massive sponges is not similarly constrained, they are expected to invest more resources into growth rather than high biomass turnover (Pawlik et al. 2016). Despite the differences in growth form, we found comparable detritus production rates by the 2 sponges, although with considerable intraspecific variability. Sponge detritus is hypothesized to be largely due to high cell turnover and shedding, particularly of sponge choanocyte cells (de Goeij et al. 2009, Alexander et al. 2014). However, sponges may also release detritus by ejecting waste products and incompletely digested food (Maldonado 2016) or via other mechanisms such as mucus production in response to sedimentation (Bell et al. 2015, Biggerstaff et al. 2017). Sponge cell turnover and shedding is reduced under suboptimal food conditions (Alexander et al. 2015b) and in wounded sponges (Alexander et al. 2015a), suggesting a complex interplay of factors such as food availability, predation, reproductive status, growth rate and sponge health may govern detritus production. Interestingly, there was higher enrichment of $^{13}$C and $^{15}$N in the sponge detritus compared with the sponge tissue (Fig. 1). We hypothesize that coral-derived C and N is preferentially incorporated into highly active cells and tissues with a high turnover rate that disproportionately contribute to sponge detritus. This would be consistent with previous studies that found that up to approximately 40% of assimilated DOM is released as detritus within 3 to 12 h, indicating rapid turnover of assimilated DOM (de Goeij et al. 2013, Rix et al. 2017).

The enrichment of $^{13}$C and $^{15}$N in the ophiuroids and polychaete confirms the last step of the sponge loop—the sponge-mediated transfer of coral-derived organic matter to higher trophic levels (Fig. 3). There are 2 possible pathways for this transfer: (1) predation on living sponge tissue or (2) uptake of sponge detritus. Spionid polychaetes (Dauer et al. 1981, Williams & McDermott 1997), as well as Ophiocoma scolopendrina and Ophiothrix ophiuroids (Magnus 1965, Warner & Woodley 1975) are well-described suspension or deposit feeders. The 2 ophiuroids were observed feeding on detritus on the sponge surface (L. Rix pers. obs.), as reported for other ophiuroid–sponge associations (Hendler 1984). Video analysis of Polydorella smurovi demonstrated characteristic spionid feeding behaviour by which particles are captured using a pair of tentaculate palps and transported to the pharynx (Naumann et al. 2016). The absence of scars or bite marks further renders direct consumption of sponge tissue unlikely; thus, we consider the enrichment of $^{13}$C and $^{15}$N in the detritivores to be due to detritus feeding. Detritus incorporation rates were comparable to those for sediment-dwelling sponge detritus feeders in the Caribbean (de Goeij et al. 2013). Combined with observations of sponge detritus feeding by sponge-associated holothuroids (Hammond & Wilkinson 1985), collectively this shows sponge detritus is utilized by a wide variety of reef fauna. Sponge detritus may be particularly important for obligate and non-motile sponge associates (e.g. Polydorella polychaetes), as continuous detritus production by sponges

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Fig. 3. The steps of the sponge loop pathway: (1) corals and algae release exudates as dissolved organic matter (DOM), (2) sponges take up DOM, (3) sponges release detrital particulate organic matter (POM), (4) sponge detritus (POM) is taken up by sponge-associated and free-living detritivores. Pathways in solid arrows indicate the steps of trophic transfer of coral-derived carbon and nitrogen demonstrated in the present study. Dashed arrows represent steps of trophic transfer from the literature: (---) Rix et al. (2017) and (---) de Goeij et al. (2013)
could help alleviate temporal fluctuations in food availability for these organisms that are reliant on their habitat to provide sufficient access to food. Nevertheless, mobile organisms also have to balance the trade-off between predator avoidance and foraging activity (e.g. Brooker et al. 2013, Catano et al. 2016), and therefore finding shelter that also supplies food can be highly advantageous (Duffy & Hay 1994). Measurable detrital C$_{\text{org}}$ and N production by the 2 sponges further supports the potential for sponge detritus to be an important resource for associated fauna, but additional studies are required to determine its quantitative importance to their diet. Food quality also influences its potential value (Andersen et al. 2007, Mitra & Flynn 2007), and less degraded detritus is typically of higher nutritional value due to selective removal of more labile fractions during degradation processes (Bowen 1987). High cell turnover and shedding is hypothesized to be the major source of sponge detritus (de Goeij et al. 2009, Alexander et al. 2014), and these freshly shed cells may be relatively undegraded. This is supported by the significantly lower C:N ratios of the sponge detritus compared with ambient suspended POM in the water column. However, sponge detritus also contains metabolic waste and incompletely digested food, which may be less labile (Maldonado 2016). Thus, compositional analysis would better establish its nutritional value. Nevertheless, we show that sponges generate food for their inhabitants—and provide them access to coral-derived energy and nutrients—as an added benefit to consider when interpreting sponge–detritivore associations. The consequences for the sponge host are less clear, but by clearing the sponge of debris, the associations may be mutualistic (Hendler 1984, Martin & Britayev 1998). However, empirical evidence of a measurable benefit (e.g. in terms of increased growth rate or reproductive output) to the sponge is lacking and could be highly context-dependent (Henkel & Pawlik 2014).

Detritivores occupy an important role in reef food webs by recycling detritus to higher trophic levels. Ophiuroids, for example, experience heavy predation, particularly by reef fish (Hendler 1984, Aronson 1988, Henkel & Pawlik 2005), thereby offering a direct pathway by which coral-derived DOM could be further transferred up the reef food web. A large fraction of reef organic matter passes through the detrital food web and these detrital pathways play an important role in recycling primary production (Alongi 1988, Hansen et al. 1992, Max et al. 2013, McMahon et al. 2016). Empirical evidence is needed to quantify the importance of the sponge loop within reef food webs, but DOM uptake by cryptic sponges is estimated to approximate gross reef primary production (de Goeij & van Duyl 2007), and trophic models suggest it may have cascading effects on energy transfer to higher trophic levels leading to altered fish production (Silveira et al. 2015). Further, the sponge loop may contribute to the efficient nutrient cycling that enables coral reefs to maintain high productivity in oligotrophic conditions (de Goeij et al. 2013). While the microbial loop also facilitates the transfer of DOM to higher trophic levels, it may rather largely fuel the pelagic food web (Worden et al. 2015), whereas transfer of DOM to higher trophic levels via the sponge loop may facilitate the recycling of reef-derived DOM to sponge-associated and other benthic fauna and thereby promote benthic productivity. Thus, this novel trophic link between corals, sponges and their associated detritivores may provide an example of how facilitative interspecific interactions not only enhance resource use between partners (Stachowicz 2001, Bruno et al. 2003), but may ultimately influence ecosystem productivity. In conclusion, there is an urgent need to recognize the pivotal role of sponges, so far largely neglected key players, within coral reef food webs.

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