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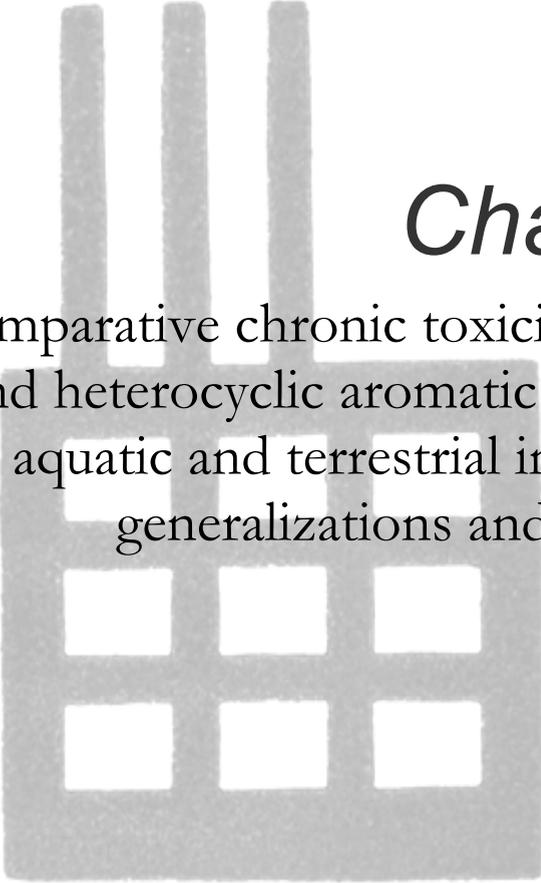
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Chapter 5

Comparative chronic toxicity of homo- and heterocyclic aromatic compounds to aquatic and terrestrial invertebrates: generalizations and exceptions.

M. León Paumen, P. de Voogt, C.A.M. van Gestel and M.H.S. Kraak. Submitted..

Abstract

The aim of the present study was to elucidate consistent patterns in chronic Polycyclic Aromatic Compound (PAC) toxicity to soil and sediment inhabiting invertebrates. Therefore we examined our experimental dataset, consisting of twenty-one chronic effect concentrations for two soil invertebrates (*Folsomia candida* and *Enchytraeus cripticus*) and two sediment invertebrates (*Lumbriculus variegatus* and *Chironomus riparius*) exposed to six PACs (two homocyclic isomers, anthracene and phenanthrene; two azaarene isomers: acridine and phenanthridine; and two azaarene transformation products, acridone and phenanthridone). In order to determine if effect concentrations were well predicted by existing toxicity- K_{ow} relationships describing narcosis, chronic pore water effect concentrations were plotted jointly against $\log K_{ow}$. Fifteen of the twenty-one effect concentrations (71%) fitted well on the acute LC50- $\log K_{ow}$ relationship, showing that narcosis was the main mode of action for the majority of the tested homo- and heterocyclic PACs during chronic exposure. Toxicity of all tested compounds to soil organisms was well described by the toxicity- K_{ow} relationship. However, for the sediment invertebrates exposed to some of the tested heterocyclic PACs deviations from narcosis (related to specific physicochemical properties of the test compounds and/or species specific sensitivities) were identified. It can be concluded that existing toxicity- K_{ow} relationships describing narcosis in some cases underestimate chronic PAC toxicity to sediment inhabiting invertebrates.

Introduction

Invertebrates inhabiting PAC contaminated soils and sediments are chronically exposed to a variety of homocyclic and heterocyclic compounds (Lahr et al., 2003; Liu et al., 2004; Uhler et al., 2005). Until now, risk assessment for PACs is based on only a limited set of homocyclic compounds, ignoring the vast number of substituted heterocyclic compounds and transformation products. This omission has gradually been recognized, and consequently heterocycles are receiving a growing scientific attention (Bleeker et al., 1999a; Feldmannova et al., 2006; Sverdrup et al., 2002a; 2002b; Wiegman et al., 2001). For some groups of heterocyclic compounds, enough data have now become available to answer the urgent question if standards for the limited set of homocyclic PACs also sufficiently protect against heterocyclic PACs.

All organic toxicants, including PACs, induce narcosis, or baseline toxicity, to some extent. Many studies have shown that narcosis is strongly related to the lipophilicity of the compound, often expressed as the n-octanol-water partition coefficient (K_{ow}) (Chen et al., 1997; De Voogt et al., 1988; Konemann, 1981; Schultz and Bearden, 1998; Swartz et al., 1995). However, specific modes of action (i.e. other than narcosis) cause deviations from such relationships (Bearden and Schultz, 1998; Escher and Hermens, 2002). Hence, the relationship between toxicity and $\log K_{ow}$ can be used in search for generalizations as well as for identifying exceptions. Bleeker et al (Bleeker et al., 2002a) applied this approach to azaarenes, heterocyclic compounds with a nitrogen substitution. In that study, acute toxicity of azaarenes to the midge *Chironomus riparius* was compared to that of homocyclic compounds by plotting the 96h LC50 values of both groups of compounds against $\log K_{ow}$. This comparison revealed that $\log K_{ow}$ was well describing the acute toxicity of both homo- and heterocyclic compounds (Bleeker et al., 2002a). This generalization is obviously a big step forward, but the same study demonstrated that specific modes of action other than narcosis caused strong deviations from the relationship between toxicity and $\log K_{ow}$ (Bleeker et al., 2002a). Such specific modes of action may even cause closely related compounds, such as isomers, to differ several orders of magnitude in toxicity (Bleeker et al., 2002a; Droge et al., 2006; Kraak et al., 1997; Leon Paumen et al., 2008a; Walton et al., 1983; Wiegman et al., 2001; Wood et al., 1983). Comparative toxicity is further complicated by the repeated observations that transformation products, generated with the aim to detoxify the parent compound, may have an unexpected and sometimes even exceptionally high toxicity. An intriguing example is the azaarene metabolite acridone, being orders of magnitude more genotoxic than its parent compound acridine and its isomer phenanthridone (Bleeker et al., 1999b).

In many short-term high-dose PAC toxicity studies, including those mentioned above, mortality was used as endpoint. This ignores that PACs can exert sublethal effects during long-term exposure, therewith affecting different biological endpoints (Droge et al., 2006; Feldmannova et al., 2006; Leon Paumen et al., 2008a). Hence, a reliable hazard assessment for PACs requires the inclusion of biological endpoints other than mortality, such as chronic effects on growth and reproduction. This motivated us to assess life cycle effects of homocyclic compounds, azaarenes and stable azaarene transformation products on soil and sediment inhabiting invertebrates (Droge et al., 2006; Leon Paumen et al., 2008a; Leon Paumen et al. 2008b). Six compounds were tested: two homocyclic compounds (anthracene and phenanthrene), two azaarenes (acridine and phenanthridine) and the two main stable transformation products of the azaarenes (acridone and phenanthridone) (Table 1). We obtained 28 d LC50 and/or EC50_{reproduction} values for the terrestrial springtail *Folsomia candida* and the enchytraeid *Enchytraeus crypticus* (Droge et al., 2006), the benthic oligochaete *Lumbriculus variegatus* (Leon Paumen et al. 2008b) and the midge *Chironomus riparius* (Leon Paumen et al., 2008a). In each of these studies some generalizations regarding chronic sublethal effects of homocyclic compounds, azaarenes and stable azaarene transformation products could be made, but many exceptions occurred, and these differed between the individual studies. Therefore, the aim of the present study was to elucidate consistent patterns in chronic PAC toxicity to soil and sediment inhabiting invertebrates. To this purpose we compared obtained chronic effect concentrations to an acute LC50-logK_{ow} relationship describing narcosis, in order to identify both the generalizations and the exceptions. The exceptions will be examined in more detail using other physicochemical properties than logK_{ow}.

Materials and methods

This study used life-cycle toxicity data obtained from studies with the terrestrial springtail *Folsomia candida* and the enchytraeid *Enchytraeus crypticus* (Droge et al., 2006), the aquatic oligochaete *Lumbriculus variegatus* (Leon Paumen et al. 2008b) and the midge *Chironomus riparius* (Leon Paumen et al. 2008a) performed earlier in our laboratories. These studies followed international guidelines (ISO, 1998; OECD, 2004a; OECD, 2004b; OECD, 2006) with slight modifications.

In short, the soil toxicity tests started with either *F. candida* that were synchronized to 10 to 12 days of age or adults of *E. crypticus* of approximately 0.4 to 0.6 cm that were adapted to clean reference soil for one day. Each replicate consisted of 30 g (wet weight) spiked soil, containing 10 individuals, in 100-ml glass jars. Test containers were kept at 20°C and 80% humidity under a 16:8-h light:dark photoperiod. The

animals were extracted from the soil and counted after the 28-day exposure period (Droge et al., 2006).

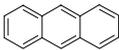
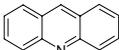
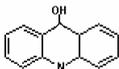
Sediment toxicity tests with the midge *C. riparius* and the oligochaete *L. variegatus* were conducted in an incubator at 20°C under a 16:8-h light:dark photoperiod. UV filters were used to suppress UV-B radiation and minimise photodegradation of the tested compounds. First instar midge larvae, less than 24 h old, were used for the toxicity tests. At the start of the experiment, ten first-instar *C. riparius* larvae were carefully introduced in each of the replicate glass beakers. Emergence was recorded daily starting at day 14, when the first adult midges started to emerge, until day 28, when the test was terminated (Leon Paumen et al., 2008a). At the beginning of the benthic oligochaete experiment, 20 *L. variegatus* adult individuals were introduced in each of the test jars. After 28 days, the sediment was sieved to extract the worms from the sediment and the number of worms in each test jar was determined, and the average reproductive output per concentration was calculated (Leon Paumen et al. 2008b).

From these life cycle toxicity tests we obtained 28-day LC50 and/or EC50_{reproduction} values for 4 test species and 6 compounds, which were used in the present study. Since for the tested compounds, having a log K_{ow} value <5, porewater exposure of the test organisms was expected (Belfroid et al., 1996), chronic pore water LC50/EC50 values (μM) were calculated from the chronic soil/sediment LC50/EC50 values ($\mu\text{mol/kg}$ soil/sediment D.W.), using experimentally determined organic carbon-water partitioning coefficients (K_{oc} values, Table 1) and the organic carbon content of the reference soil (2.2%) and the reference sediment (8.1%). The calculated porewater LC50 and EC50_{reproduction} values were plotted against the log K_{ow} values of the tested compounds.

Results and Discussion

Our experimental matrix consisted of six compounds tested on four invertebrates, with one (sub)lethal endpoint for the aquatic organisms and two endpoints, survival and reproduction, for the terrestrial organisms, resulting in thirty-six potential effect concentrations. In twenty-one of the thirty-six cases a reliable effect concentration was obtained (Table 2), in all other cases no toxic effects were observed. We plotted these chronic LC50 (Figure 1A) and EC50 (Figure 1B) values on the relationship between acute aquatic LC50 data and log K_{ow} by Bleeker et al. (Bleeker et al., 2002a), which describes narcosis. Figure 1 shows that all possible outcomes were actually ob-

Table 1. Polycyclic Aromatic Compounds (PACs) selected for this study and some of their properties: molecular weight (MW), octanol-water partitioning coefficient ($\log K_{OW}$), soil/sediment-water partitioning coefficient ($\log K_{OC}$), and water solubility (Sw).

Compound	Structure	MW	$\log K_{OW}$	$\log K_{OC}$	Sw (μM)
Homocyclic PACs					
anthracene		178.23	4.53 ^a	4.3 ^b	0.37 ^c
phenanthrene		178.23	4.48 ^a	4.2 ^b	7.20 ^c
Heterocyclic PACs (azaarenes)					
acridine		179.22	3.27 ^a	4.1 ^b	260 ^c
phenanthridine		179.22	3.44 ^a	3.8 ^b	25.7 ^d
Azaarene transformation products					
acridone (enol)		195.22	2.95 ^e	3.9 ^b	670 ^d
phenanthridone (keto)		195.22	2.70 ^f	4.3 ^b	1428 ^d

^a Experimental values from (Helweg et al., 1997).

^b Values from (Jonassen et al., 2003).

^c Values from (Pearlman et al., 1984).

^d Calculated with WSKOW, version 1.40.

^e Experimental values from (Thomsen, 2002).

^f Values from (Bleeker et al., 2002b).

served: some compounds were less toxic (e.g. anthracene and phenanthridone) while others were more toxic than expected (e.g. phenanthridine). To aid the discussion on generalizations and exceptions, the results of the comparisons made in Figure 1 are summarized in Table 3. Generalizations will be discussed first, followed by the five different types of exceptions that were found.

1. Generalizations

The main purpose of the present study was to answer the question if chronic effect concentrations for homo- and heterocyclic compounds and transformation products were well predicted by existing toxicity- K_{ow} relationships. Fifteen of the twenty-one effect concentrations (71%) fitted well on the acute LC50- $\log K_{ow}$ relationship (Figure 1 and '=' marks in Table 3), meaning that narcosis was the main effect for the majority of the tested homo- and heterocyclic compounds during chronic exposure. Nevertheless, phenanthrene was the only compound for which nearly all effect concentrations agreed well with the acute LC50- $\log K_{ow}$ relationship, except for the LC50 for *E. crypticus*, which was higher than expected (Table 3). The homocyclic compound phenanthrene is a well-known narcotic compound, and the results from this study are in agreement with the available literature (Landrum et al., 1992; Neilson, 1998; Sverdrup et al., 2002b). For the two heterocyclic compounds, the eight effect concentrations for the two terrestrial invertebrates agreed well with the acute LC50- $\log K_{ow}$ relationship, except for the phenanthridine LC50 for *E. crypticus*, which was higher than expected (Table 3). Hence, it can be concluded that chronic soil toxicity for homo- and heterocyclic PACs could be well explained by an acute effect- $\log K_{ow}$ relationship describing narcosis, as suggested in previous studies (Sverdrup et al., 2002b).

For the transformation products only two effect concentrations were obtained, one agreeing with the relationship (acridone LC50 for *C. riparius*) and one being lower than expected (acridone EC50 for *L. variegatus*) (Table 3). Due to the limited number of effect concentrations (only 2 out of 12 potential effect concentrations) a reliable comparison with toxicity of their heterocyclic parent compounds and homocyclic analogues can not be made. On the other hand, this result shows that PAC metabolism generally results in transformation products that are harmless for most organisms, although unexpected high metabolite toxicity may be observed occasionally (see below).

2. Exceptions

2.1. Compounds for which the predicted effect concentration was above maximal water solubility

For the homocyclic compound anthracene the predicted effect concentration (0.62 μM , Table 3) is above maximal water solubility (0.37 μM , Table 1). Hence, no adverse effects of anthracene at maximal water solubility were expected, which was indeed the case for the two terrestrial invertebrates (Table 3). In the experiments with the midge

Table 2. Calculated pore water effect concentrations (μM) for the effect of six Polycyclic Aromatic Compounds (PACs) on survival (LC50) and reproduction (EC50) of terrestrial and benthic invertebrates (*Folsomia candida*, *Enchytraeus crypticus*, *Chironomus riparius* and *Lumbriculus variegatus*).

	log K_{ow}	SOIL						SEDIMENT	
		<i>F. candida</i>		<i>E. crypticus</i>		<i>C. riparius</i>	<i>L. variegatus</i>		
		LC50	EC50	LC50	EC50	LC50	LC50	EC50	
Homoc.									
	4.53	-	-	-	-	-	-	0.3	
	4.48	1.08	0.8	10.9	1.1	0.5	0.4		
Azar.	3.27	14.9	9.8	18.3	15.4	1.4	1.2		
	3.44	8.9	3.7	23.9	7.3	0.9	0.6		
TR. pr.	2.95	-	-	-	-	4.2	0.5		
	2.70	-	-	-	-	-	-		

Table 3. Summary of the comparison between chronic effect concentrations for soil and sediment invertebrates and acute LC50s (narcosis) from (Bleeker et al., 2002b). **w**: not toxic at highest tested concentration, which was above water solubility; **c** not toxic at highest tested concentration, which was below the acute LC50; **n** not toxic at the highest tested concentration, which was above the acute LC50; **-** chronic effect concentration higher than acute LC50, = chronic effect concentration similar to acute LC50 (narcosis), **+** chronic effect concentration lower than acute LC50 (specific effects besides narcosis).

	log K_{ow}	acute LC50 (μM , Bleeker et al. 2002)	SOIL						SEDIMENT	
			<i>F. candida</i>		LC50/EC50	<i>E. crypticus</i>		LC50/EC50	<i>C. riparius</i>	<i>L. variegatus</i>
			LC50	EC50		LC50	EC50			
Homoc. anthracene	4.53	0.62	w	w		w	w		c	=
Homoc. phenanthrene	4.2	1.16	=	=	1.4	-	=	3.8	=	=
Azar. acridine	3.27	6.67	=	=	1.6	=	=	1.5	=	+
Azar. phenanthridine	3.44	4.85	=	=	3.6	-	=	3.0	+	+
Trans. pr. acridone	2.95	12.18	n	n		n	n		=	+
Trans. pr. phenanthridone	2.7	19.5	c	c		c	c		c	c

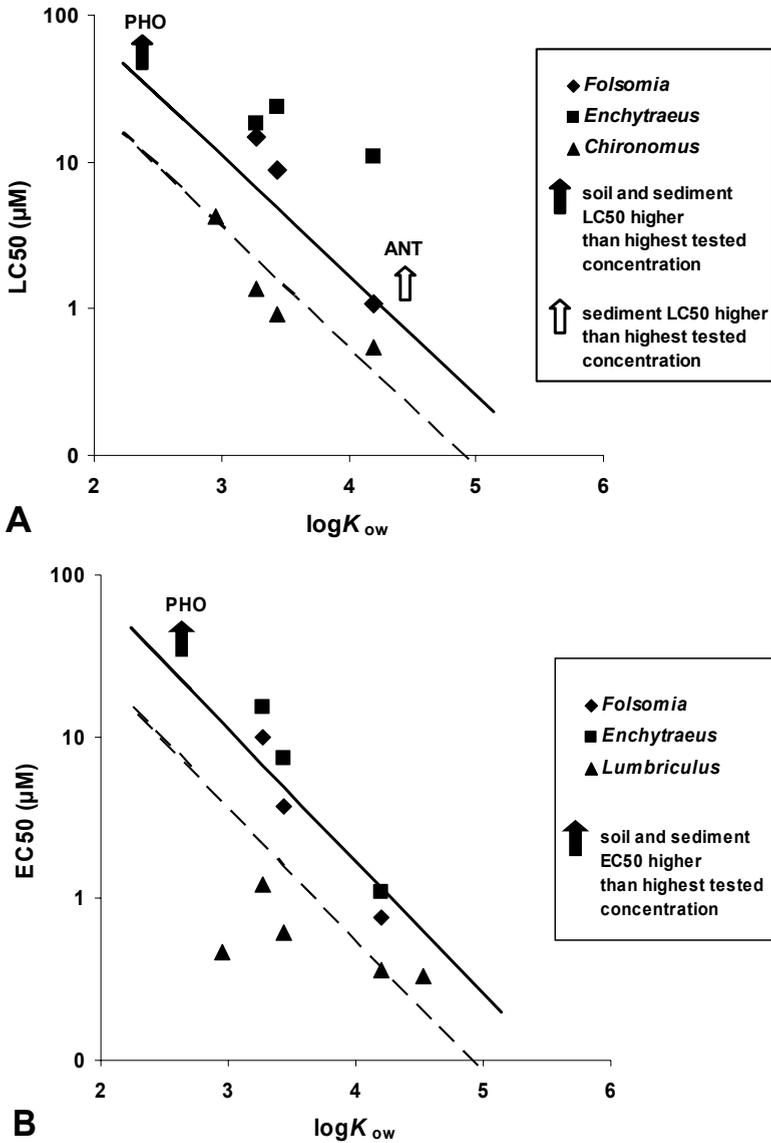


Figure 1. Effect concentrations for the effect of six Polycyclic Aromatic Compounds (PACs) on survival (LC50, **A**) and reproduction (EC50, **B**) of terrestrial and benthic invertebrates (◆ *Folsomia candida*, ■ *Enchytraeus crypticus*, ▲ *Chironomus riparius* / *Lumbriculus variegatus*), as a function of logK_{ow}. Solid line: linear logLC50-logK_{ow} relationship for *Chironomus riparius* exposed for 96 hours to similar PACs ($y = -0.8162x + 3.4936$, $r^2=0.986$, (Bleeker et al. 2002). Dashed line: ratio LC50 from (Bleeker et al. 2002) / 3, lower limit for narcosis.

C. riparius, anthracene was dissolved in acetone and added to wet sediment, but due to its low water solubility unreliable spiking results were obtained at higher anthracene concentrations in the sediment (Leon Paumen et al., 2008a). Therefore, the highest reliable test concentration (0.15 μM) was too low to observe any adverse effects. In contrast, in the experiments with the benthic oligochaete *L. variegatus* a different spiking method using dry sediment was applied, and reproduction was affected at a concentration (0.33 μM , (Leon Paumen et al., 2008a)) similar to the maximal water solubility (0.37 μM). The similarity between expected effect concentrations and water solubility makes the toxicity of anthracene rather unpredictable: due to species specific sensitivities, for some species effect concentrations may be above maximal water solubility and no effects will be observed. In contrast, for other species (*L. variegatus* in the present study) effect concentrations are below maximal water solubility, and because of high hydrophobicity of anthracene this results in the lowest effect concentration of the six tested compounds. Likewise, low reproduction EC10 values for anthracene compared to EC10s for other PACs were observed for the springtail *F. fimetaria* (Sverdrup et al., 2001), while the model used in that study predicted that no chronic toxicity of anthracene would be observed due its low water solubility.

2.2. Compounds which were not toxic at the highest tested concentration, which was below the predicted effect concentration

The two transformation products used in this study, acridone and phenanthridone, can occur in two tautomeric forms (keto and enol). In the present experiments, where exposure occurs via the water phase, the most water soluble tautomer is probably the form of the compound that exerts a toxic effect, as stated in Bleeker et al. 1999 (Bleeker et al., 1999b). Thus, phenanthridone would be mainly in keto form, having the lowest lipophilicity (K_{ow}) and the highest water solubility (S_w) of all tested compounds (Table 1). Therefore, expected effect concentrations were highest (Table 3). Because of the low solubility of phenanthridone in acetone, the high expected pore water effect concentration was not achieved with the spiking method used, and phenanthridone did not exert adverse effects on survival, emergence or reproduction of any of the test organisms (Table 3).

2.3. Compounds which were not toxic at the highest tested concentration, although this concentration was above the predicted effect concentration

The transformation product acridone did not affect survival and reproduction of the terrestrial invertebrates, although the highest tested pore water concentration was far above the acute LC50. Hence, effects of acridone were expected, but not observed. In contrast, acridone affected survival or reproduction of the two aquatic

invertebrates at or even below expected concentrations (Table 3). Since narcosis is the minimal toxicity that a compound exerts, explanations have to be found for the lack of narcotic effects of acridone to the two terrestrial invertebrates. One reason could be the experimental K_{oc} value used for the calculation of pore water concentrations (Jonassen et al., 2003, (Table 1), which was much lower than the value for its isomer phenanthridone. This, together with the low organic carbon content of the soil, could lead to an overestimation of the pore water concentration in the soil. Another explanation could be the faster depletion in the soil compared to water-saturated sediment, which could limit the availability of the compounds to the test organisms (Jager et al., 2000). These considerations underline the fundamental problems in research on scarcely studied heterocyclic PAC and PAC transformation products: although there is an urgent need for insight in their fate, effects and risk, it is still hard to find reliable values for their physicochemical properties and accurately estimate their availability in soil and sediment.

2. 4. *Compounds for which the observed effect concentration was above the predicted effect concentration*

Two of the 21 observed effect concentrations were above the predicted effect concentration, i.e. effect concentrations for the homocyclic compound phenanthrene and its azaarene analogue phenanthridine on survival of the terrestrial oligochaete *E. crypticus*. *E. crypticus* is known for its low sensitivity to organic compounds compared to the springtail *F. candida* (Droge et al., 2006; Kolar et al., 2008; Krogh et al., 2007; Sverdrup et al., 2002d), while toxic effects of heavy metals occur at the same range of concentrations for these two soil invertebrates (Kuperman et al., 2006; Kuperman et al., 2004; Lock and Janssen, 2003).

Toxic effects of narcotic compounds are mainly caused by unspecific partitioning of the toxicant into lipids in cellular membranes, leading to membrane instability. Cellular membrane composition is quite stable across different groups of organisms. For this reason, the Critical Body Residue concept (CBR, (McCarty and Mackay, 1993)) states that non polar narcotics (e.g. PACs, PCBs) cause 50% mortality at a relatively narrow body concentration interval expressed on lipid basis. This concept has been successfully applied to interpret toxicity data for sediment and soil inhabiting organisms (Fisher et al., 1999; Lee et al., 2002; Leon Paumen et al., 2008b; Leslie et al., 2004b). Because of the unspecific character of the narcotic mode of action, it has been suggested that this threshold would be applicable to all organisms. The present study demonstrates, however, that in spite of the narrow range of critical body residues, species specific variation does occur, leading to deviations from existing $LC50 - \log K_{ow}$ relationships and complicating effect prediction.

2. 5. Compounds which were more toxic than predicted

Four of the twenty-one observed effect concentrations (19%) were lower than expected (Table 3). Since these were all sediment effect concentrations for heterocyclic compounds, this means that for benthic organisms, toxicity of heterocyclic PACs is underestimated using effect- K_{ow} relationships describing narcosis. Three out of the four exceptions were observed for reproduction of the oligochaete *L. variegatus* (EC50s for acridine, phenanthridine and acridone), and two out of the four exceptions were observed for the two benthic invertebrates exposed to phenanthridine.

The high sensitivity of *L. variegatus* reproduction to three of the four tested heterocyclic PACs could be related to its asexual mode of reproduction. The soil organisms used in this study and the midge *C. riparius* are oviparous, but the oligochaete *L. variegatus* reproduces via asexual fragmentation (architomy) under laboratory conditions. In the architomic fission process two fragments are formed and the missing segments are generated during a regeneration period that lasts more than a week (Leppanen and Kukkonen, 1998; Martinez et al., 2005). During this period the newly generated worms do not feed, and high sensitivity to toxicants in the sediment might be related to the high energy demands for segment regeneration, which could influence energy allocation to detoxification mechanisms (Penttinen and Kukkonen, 1998).

The high toxicity observed for the benthic organisms exposed to phenanthridine compared to its isomer acridine illustrates the drawback of using K_{ow} as a descriptor in studies dealing with isomers. K_{ow} is a macroscopic property, and therefore differences between K_{ow} values of isomers will be minimal. Hence, other more specific physicochemical properties are needed to explain toxicity differences between closely related compounds. For instance, higher toxicity of phenanthridine compared to other benzoquinolines on mussel filtration rate has been related to a combination of physicochemical properties describing the attractive and repulsive forces governing toxicant-membrane interaction (Kraak et al., 1997). Following this approach, several physicochemical properties of the six compounds analyzed in this study were calculated using ChemProgPro™ and ClogP™ (e.g. charge of the N atom, charge of the O atom, dipole, HOMO-LUMO gap). For the tested azaarene isomers the dipole of the molecule, which defines its electronic conformation, differed by a factor of 2.5, and could be the reason for the high toxicity of phenanthridine compared to acridine. The differences between the electronic conformation of the two isomers have obvious implications in toxicant-membrane interaction and receptor binding.

However, at the moment too little information is available to link observed effects of PACs to physicochemical properties other than K_{ow} and in a quantitative way.

Conclusions

Narcosis was the main mode of action of the tested homo- and heterocyclic PACs during chronic exposure of soil and sediment invertebrates. However, exceptions related to specific physicochemical properties of the compounds and/or species specific sensitivities were also identified. Particularly benthic invertebrates were sometimes more sensitive to the tested heterocyclic PACs than expected, meaning that PAC sediment risk assessment based solely on a small set of homocyclic structures could be underprotective.